

# Europe's other debt crisis caused by the long legacy of future extinctions

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**Rapid economic development in the past century has translated into severe pressures on species survival as a result of increasing land-use change, environmental pollution, and the spread of invasive alien species. However, though the impact of these pressures on biodiversity is substantial, it could be seriously underestimated if population declines of plants and animals lag behind contemporary environmental degradation. Here, we test for such a delay in impact by relating numbers of threatened species appearing on national red lists to historical and contemporary levels of socioeconomic pressures. Across 22 European countries, the proportions of vascular plants, bryophytes, mammals, reptiles, dragonflies, and grasshoppers facing medium-to-high extinction risks are more closely matched to indicators of socioeconomic pressures (i.e., human population density, per capita gross domestic product, and a measure of land use intensity) from the early or mid-, rather than the late, 20th century. We conclude that, irrespective of recent conservation actions, large-scale risks to biodiversity lag considerably behind contemporary levels of socioeconomic pressures. The negative impact of human activities on current biodiversity will not become fully realized until several decades into the future. Mitigating extinction risks might be an even greater challenge if temporal delays mean many threatened species might already be destined toward extinction.**

extinction debt | socioeconomic history | time lag

The progressive impact of environmental degradation on the loss of global biodiversity (1–4) is strongly linked to key socioeconomic indicators such as human population size (5), land use (6), and gross domestic product (GDP) (7, 8). However, species populations do not necessarily respond immediately to environmental degradation but might do so with a delay (9, 10). Such time-lags between environmental forcing, population decline, and, finally, extinction create a transient disequilibrium between environmental conditions and species' abundance or range size, which has been conceptualized as "extinction debt" (9). Recent empirical research has shown that, at the scale of individual habitat patches, a delayed response of species to habitat loss and fragmentation is indeed often detectable, particularly among habitat specialists (9–11). The likelihood and magnitude of extinction debt is still contentious, however (12, 13), and seems to vary with the nature of environmental degradation and with the life history traits of the species concerned (14–18). In long-lived or less-mobile taxa (e.g., vascular plants, bryophytes, reptiles), a delayed response of populations to the deterioration and fragmentation of their habitats is especially likely and might extend over at least several decades.

If time-lags in population decline at the scale of individual habitats are a common phenomenon, the number of species facing extinction risks at larger spatial scales might easily be underestimated because many local populations of extant species might not survive in the long-term even if further environmental degradation is halted. Thus, red lists describing threatened species at

either a national or global scale, which identify species extinction risks using criteria such as a reduction in population size, range size, and perceived rate of recent population or range decline (19), might be too optimistic. However, despite the recent increase in interest in the potential for temporal lags in population declines and extinction processes (8, 10, 12, 17), evidence for extinction debt has not yet been examined for a broad range of taxa at a spatial scale consistent with national or global decision-making. Given that many conservation policies in Europe (and elsewhere) are developed and implemented at regional to global scales (e.g., Ramsar Convention and Convention on Biological Diversity) and reported at the level of individual countries, assessing extinction debt at this scale would appear to be a crucial step.

Here, we provide such an assessment by analyzing national red lists (NRL) of threatened species of seven taxonomic groups (vascular plants, bryophytes, mammals, fish, reptiles, dragonflies, and grasshoppers) from 22 European countries (Tables S1 and S2) in relation to the magnitude of contemporary as well as historic drivers of environmental degradation. Red lists provide an objective framework for the classification of the broadest range of species according to their extinction risk based on the current trend in population size, geographic range, and area of occupancy (19). We focus on Europe because this is the only continent where NRLs have become recently available, and are regularly updated, for a range of taxonomic groups across more than 20 countries using standard approaches to the listing of threatened species. Our analysis is based on a rationale similar to the one used in Essl et al. (5) for evaluating delays in alien species invasions: if NRLs provide an accurate estimate of the number of species at risk under the current magnitude of human pressures on biodiversity, these numbers should best correlate with the spatial variation in contemporary levels of these drivers. If, by contrast, time-lags between different human pressures and subsequent species responses occur, the number of threatened species should better reflect historic rather than contemporary values of these drivers. We measure the values of these drivers by three indicators of human socioeconomic activities that are known proxies of environmental pressures on biodiversity (2, 7, 8, 20): human population density (PD), per capita GDP, and human appropriation of net primary production (HANPP) (21). We estimated HANPP as the ratio of the net

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**Table 2. Proportions of threatened species from seven taxonomic groups and 22 European countries as explained by single variable models of historical or current socioeconomic indicators**

	1900	1950	2000	1900	1950	2000	1900	1950	2000
PD	0.18	0.15	0.13						
GDP				0.23	0.14	0.01			
HANPP							0.35	0.23	0.16
$R^2_{MF}$	0.06	0.04	0.03	0.10	0.04	0.00	0.18	0.08	0.04
AIC	5,837	5,972	6,043	5,599	5,986	6,213	5,085	5,730	5,950
AW	>99.9	<0.01	<0.01	>99.9	<0.01	<0.01	>99.9	<0.01	<0.01

Fixed-effect estimates, McFadden's  $R^2$ , Akaike information criterion (AIC), and Akaike weights (AW) for generalized linear mixed-effects models relating the variation in the proportion of red-listed species to either historical (1900 and 1950) or contemporary (2000) values of impact indicators. All fixed-effect estimates are different from 0 with a probability >0.99.

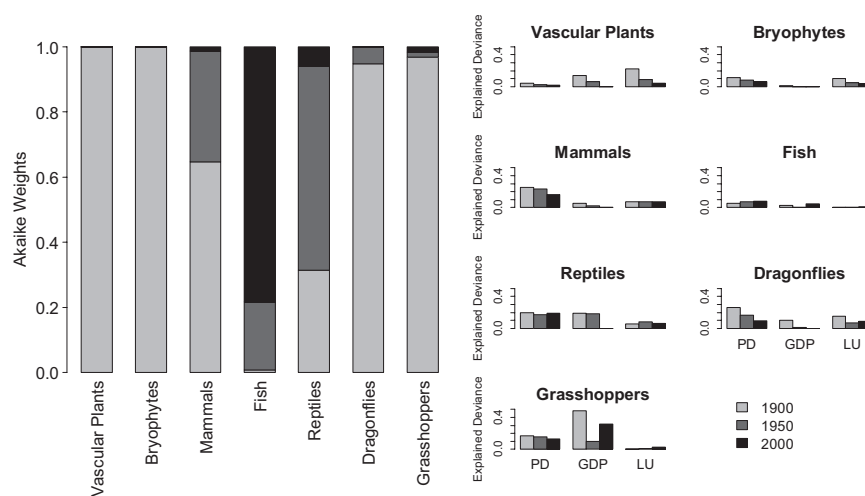
expenditures as an additional independent variable. This extension indeed improved model fit (compare Tables 1 and 2 and Table S3; Akaike weights >99.9 in favor of the models with the expenditures included); however, it did not change the relative ranks of the models from 1900, 1950, and 2000, except for fish, which appear most closely correlated to socioeconomic activity from 1950 rather than 2000, when expenditures are taken into account (Fig. S2).

### Discussion

Even though human impacts on the environment greatly intensified following World War II (3, 6, 23, 24), our results suggest that the current threat status of many species reflects a legacy of several decades, and for four taxa even as much as a century. Such extended time-lags have indeed been observed in habitat-scale studies on vascular plants (12, 18) or cryptogams (25) for which historic indicators most markedly outperformed current indicators in explaining the proportion of threatened species in our country-scale study as well (Table 3). Contrary to common expectations (10, 26), however, the risks faced by taxonomic groups represented by short-lived species such as dragonflies and grasshoppers seem to reflect human impacts on the environment with a delay similar to that for plants. For a similar taxonomic group, butterflies, the available empirical evidence for delayed population response to habitat loss is mixed with some studies reporting rather fast

(several years) (12, 27) and others relatively long (several decades or more) relaxation times (17). In addition, theoretical simulations have suggested that the delay in insect metapopulation extinctions can extend well beyond 100 y (28), and that their lag times are particularly long if available habitat networks are reduced to levels close to the extinction threshold (29), as might be the case in many landscapes of Europe where an intensively used agricultural matrix is interspersed by remnants of near-natural and nonintensively used habitats (30).

For the vertebrate taxa, time-lags between socioeconomic indicators and population decline appear to be less pronounced. In particular, fish were the only taxonomic group for which extinction risks appear more closely correlated to contemporary indicators. We do not know why fish behave differently, but it might be that anthropogenic impacts on freshwater ecosystems, such as water pollution, channelization, construction of dams, and water abstraction (31), have a more immediate effect because they not only reduce the quality and quantity of habitats, but directly and uniformly modify the medium in which species live. Additionally, historic legacies might be masked by the particularly intense but more recent modification of the aquatic environments (23). Mammals, however, might be particularly sensitive to habitat loss and fragmentation because they often need large contiguous habitats for survival (32) and can hence reach their extinction thresholds (29) during an earlier stage of habitat degradation and loss.



**Fig. 2.** Proportions of species facing medium-to-high extinction risks from seven taxonomic groups in 22 European countries (for data sources, see Table S2) as explained by current and historic socioeconomic models. (Left) Relative support (Akaike weights) for multiple logistic regressions models explaining the proportions of red-listed species by historical (1900, light gray; 1950, dark gray) or contemporary (2000, black) levels of socioeconomic impact indicators. The three indicators (human population density, per capita GDP, human appropriation of net primary productivity) were subjected to a principal component analysis before fitting the regression models to remove multicollinearity among explanatory variables. (Right) The deviance explained by simple logistic regression models relating the proportion of threatened species separately to either historical or contemporary levels of the three individual indicators: GDP, per capita GDP; LU, land use, i.e., human appropriation of net primary productivity; and PD, human population density.

**Table 3. Goodness of fit of models relating the proportion of threatened species from seven different taxonomic groups across 22 European countries to either historical or contemporary socioeconomic impact indicators**

	AIC <sub>1900</sub>	AIC <sub>1950</sub>	AIC <sub>2000</sub>	D <sup>2</sup> <sub>1900</sub>	D <sup>2</sup> <sub>1950</sub>	D <sup>2</sup> <sub>2000</sub>
Vascular plants	3,589	4,458	4,824	0.30	0.13	0.05
Bryophytes	828	861	873	0.14	0.10	0.09
Mammals	177	179	185	0.27	0.25	0.19
Fish	212	206	203	0.01	0.15	0.17
Reptiles	139	138	143	0.26	0.27	0.22
Dragonflies	171	177	193	0.26	0.22	0.10
Grasshoppers	103	111	111	0.51	0.39	0.40

Akaike Information Criterion (AIC), and Explained Deviance (D<sup>2</sup>) of multiple logistic regression models. The three indicators (human population density, per capita GDP, and human appropriation of net primary productivity) were subjected to a principal component analysis before fitting the regression models to remove multicollinearity among explanatory variables.

In conclusion, our results suggest that our perception of extinction risks at large geographical scales lags considerably behind increases in socioeconomic pressures on biodiversity across different taxonomic groups. With respect to the future, this long lag time implies that contemporary human pressures have already influenced the threat that biodiversity will face in the next decades, and this is underestimated by current NRLs. Indeed, the rapid recent increase in socioeconomic activity has triggered new, or at least greatly intensified, pressures on biodiversity arising from processes such as atmospheric nitrogen or acidic deposition (33), biological invasions (5, 34, 35), and climate change (2), which now reinforce and synergistically interact with habitat loss and fragmentation, the main drivers of population decline and extinction in the past. Although these novel pressures have rarely been studied in an extinction debt context (10), their effects on biodiversity might be characterized by similarly long lag times (5, 18). For example, both recent observations (36) and modeling studies (37) suggest that range adaptations of native species to changing climates considerably lag behind the velocity of climate change (38) and that remnant populations (39) currently occupy sites that are no longer climatically suitable for them in the long run (18).

Given this accumulating extinction debt from different sources, reducing further pressures on ecosystems might not be sufficient to reduce future biodiversity loss. In addition, long lag times might also blur cause–effect relationships, and if delayed responses by species to historic drivers (e.g., land use change) are confused with responses to more recent phenomena (e.g., climate change), mitigation measures might target the wrong drivers. Our results, combined with the evidence of both global (4) and European (40) failure to reach the 2010 biodiversity target, suggest current commitments to stop biodiversity loss in the region are even more inadequate than currently appreciated. From a global perspective, the extended time lags in extinction risk in Europe might also be a feature of other industrialized regions that have followed a similar development path—e.g., North America and Australia (6). Future measures to counter extinctions in these areas will probably require the mobilization of efforts well beyond current investment in conservation policies and importantly account for both new drivers of population declines as well as for those that have acted for over a century. Minimizing the magnitude of the “sixth extinction crisis” (41) might be an even greater challenge when temporal delays are taken into account.

## Materials and Methods

**Total and Red List Species Numbers.** We extracted total species numbers from national checklists, standard faunas, and floras, with some updates by national experts. Numbers of threatened species included on red lists were taken from the most recent NRLs, >90% of which have been published

between 1995 and 2010 (Table S2). We have included only species that face medium-to-high extinction risks (IUCN categories EN, VU, and CR) (19), which are generally referred to as threatened species. We excluded species that had already gone extinct [extinct (EX) and extinct in the wild (EW)] to avoid bias in favor of an historical explanation. We used the recent NRLs in Europe that were developed following standardized criteria established for national to global red listing led by the IUCN since the 1990s (19). The five IUCN criteria for assessing extinction risks are based on current status and trends in population size, geographic range, and area of occupancy, and include quantitative analyses of extinction risks (e.g., population viability analysis).

**Socioeconomic Indicators.** The scarcity of historical socioeconomic data limited our sample size to 22 countries (Table S1), but these are representative of the variation in European socioeconomic trajectories and include nations from both sides of the former Iron Curtain. Data on current and historical population densities and per capita GDP were taken from the Total Economy Database ([www.ggd.net/databases/ted.htm](http://www.ggd.net/databases/ted.htm)). We standardized current and historical per capita GDP to 1990 International (Geary–Khamis) dollars, a hypothetical currency unit with the same purchasing power that the US dollar had in the United States in 1990. In calculating net primary productivity harvested by humans NPP<sub>h</sub>, we used a broad definition of harvested NPP (21), which encompasses biomass extracted for further economic use and biomass destroyed during harvest.

**Calculation of HANPP.** As a proxy for land use intensity we used an indicator that measures the human impact on trophic energy availability in ecosystems, thereby quantifying the human influence on certain important ecosystem processes (42, 43). We divided the total amount of harvested NPP<sub>h</sub> by the NPP of the potential vegetation, i.e., the vegetation assumed to exist in the absence of human land use (NPP<sub>0</sub>) to take biogeographic variation in natural productivity into account. The ratio NPP<sub>h</sub>/NPP<sub>0</sub> is an indicator of land use intensity, and it is one variant of HANPP for which several empirical studies (44–46) support the hypothesis that it is a valid indicator of pressures on biodiversity (47). Different authors have used different definitions of HANPP. Vitousek et al. (42) proposed three definitions; the definition used here is similar, but somewhat less inclusive than Vitousek’s intermediate definition also adopted by Rojstaczer et al. (48) and Imhoff et al. (49). Vitousek’s intermediate definition includes the entire NPP of intensively managed ecosystems, whereas NPP<sub>h</sub> includes only plants actually harvested or destroyed during harvest.

We use an encompassing concept of biomass harvest that also includes biomass grazed by livestock and biomass destroyed during harvest, such as belowground biomass on cropland (21, 43). Data on the harvest of crops and timber are derived from statistical sources (50–54). Grazed biomass is estimated on the basis of feed balances using livestock and market feed data from the Food and Agriculture Organization and demand factors that considered species, region, and time (21, 55). Extraction of crop residues, harvest losses in forestry, and belowground biomass on cropland and harvested forest areas are estimated using region and time-dependent coefficients based on Krausmann et al. (56). NPP<sub>0</sub> was calculated for 1910 and 2000 (Table S4) with the LPJmL model (57). Both NPP<sub>h</sub> and NPP<sub>0</sub> are measured in terms of carbon content.

For a few countries, historic land-use data were only available for 1910. However, because land-use changes in Europe primarily occurred in the post-World War II period (6), we consider this approach to be justified.

**Data Analysis.** National floras and faunas of European countries vary considerable in species richness due to differences in country area, climatic conditions, biogeographic setting, and other factors. We accounted for this uneven distribution of regional biodiversity by considering the numbers of species on NRLs as proportions of the countries’ total native floras and faunas (Table S2): our response variable was hence the average probability of a species from a country-specific species pool to be on a particular country’s red list in the early 21st century. For analyzing the data across all seven taxonomic groups, we related the proportions of threatened species with the three socioeconomic variables by means of generalized linear mixed-effects models (GLMMs) with a logit link function for binomial error distributions (function *glmer* of the R package *lme4*) (58). We used each of the seven taxonomic groups as a grouping variable and estimated random effects for the intercept. More complex model structures with random effects for each predictor variable did not converge with the available fitting algorithms, especially in the case of multivariate regression models (see below).

For contemporary and historical economic conditions, we fitted such models separately for each of the three socioeconomic variables and for all three of



them together. In the latter case, we first subjected the three variables of contemporary and historic socioeconomic pressures to principal component analyses (PCA) separately for each of the time points and then used the scores of the 22 countries on the three axes of these PCAs as explanatory variables in the GLMMs to avoid problems possibly arising from multicollinearity. The PCAs were done by means of a singular value decomposition of the centered and scaled data matrices as implemented in the R function *prcomp* (59).

For each fitted GLMM, we calculated the corrected Akaike information criterion ( $AIC_c$ ), which includes a second-order bias correction appropriate for small sample sizes (60). We then compared the  $AIC_c$  of each model pair [with the same independent variable(s) from 1900, 1950, or 2000] by means of Akaike weights. Akaike weights are derived from  $AIC_c$ s and quantify the relative support for individual candidate models as the probability that each of them delivers the best explanation for a given dataset (60). In addition, we assessed goodness-of-model fit by calculating McFadden's  $R^2$  values (61), defining the respective null models as GLMMs with an intercept term as the only fixed effect (and the same random effects).

Regression models that include each of the socioeconomic variables separately or in combination were also fit for each taxon separately using logistic regression [generalized linear models (GLMs) with a logit link function] instead of GLMMs. Goodness-of-GLM fit was evaluated by calculating the explained deviance (62) ( $D^2$ ), and multivariable single-taxon models were compared by means of Akaike weights in the same way as described for GLMMs. Using the geographical coordinates of the country's capitals, we checked the residuals of all GLMs for spatial autocorrelation by calculating Moran's  $I$  for a neighborhood radius of 1,000 km (to guarantee that each country has at least one neighbor in the neighborhood matrix). The probability of getting an  $I$  value higher than the empirical one purely at random was assessed by means of 999 permutations of the vector of model residuals using the function *moran.mc* implemented in the R package *spdep* (63). To safeguard against possible bias from spatial autocorrelation, we refitted

GLMs with a low such probability ( $<0.1$ ) as autologistic models, i.e., by additionally introducing an autocovariate calculated by means of the function *autocov\_dist* in the R library *spdep* (63). Residual autocorrelation was reduced in these autologistic models (with the exception of the grasshopper model for the year 2000, Table S5), and the probability of randomly generating an higher residual autocorrelation hence increased, but in no case was the ranking of goodness-of-fit reversed between historic and contemporary models compared with the respective ordinary GLMs. We hence consider our results robust and qualitatively unaffected by spatial autocorrelation.

To test for an eventual bias due to recent investments into environmental protection, we collected data on recent levels of public environmental expenditures of the 22 countries (in percent of GDP) extracted from the European Commission's Eurostat Web site (<http://epp.eurostat.ec.europa.eu/tgm/table.do?tab=table&init=1&language=en&pcode=ten00049&plugin=0>, accessed May 10, 2011; Table S4). We calculated the mean of the available data from the years 1998–2009, and repeated all multivariable GLMMs and GLMs using these expenditures as an additional covariate. All statistical analyses were done in R 2.13.1 (60).

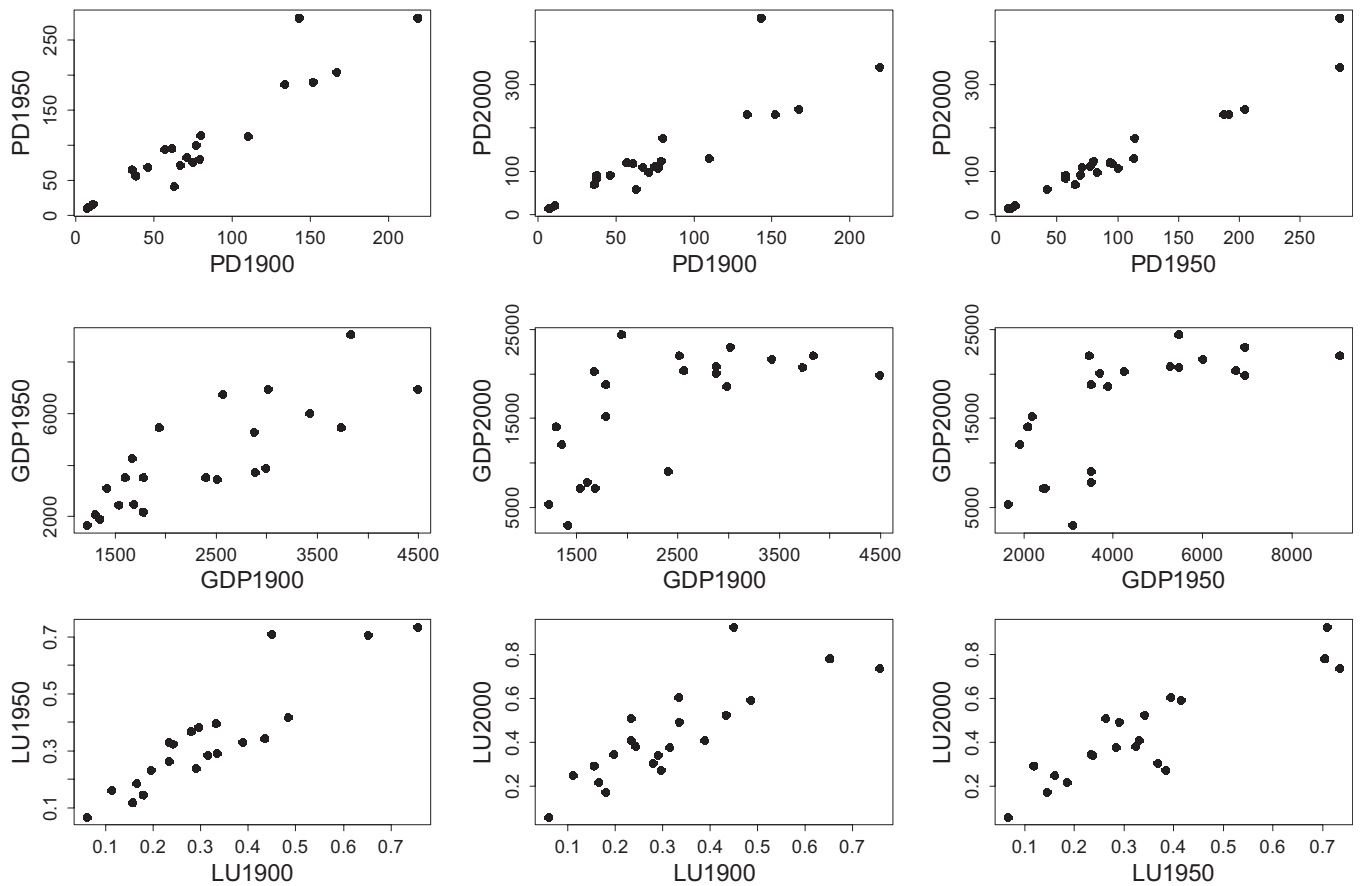
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# Supporting Information

Dullinger et al. 10.1073/pnas.1216303110



**Fig. S1.** Temporal correlations among socioeconomic variable levels in 22 European countries. GDP, per capita domestic product in 1,000 International Geary-Khamis follars; LU, land use, i.e., human appropriation of net primary productivity (NPP) as a proportion of the net productivity of the potential vegetation (NPP<sub>0</sub>); PD, human population density in people per square kilometer.











Table S2. Cont.

References	Country	Taxon	Data type
Smith AJE (2003) Chapter 4: Mosses, liverworts and hornworts. In Hawksowrth DL (ed) The changing wildlife of Great Britain and Ireland. (CRC Press).	Ireland and United Kingdom	Bryophytes	RL, CL
Bulgarini F, Calvario E, Fraticelli F, Petretti F, Sarrocco S (2007) Libro Rosso degli animali d'Italia. Vertebrati. WWF Italia.	Italy	Vertebrates	RL, CL
Aleffi M, Schumacker R (1995) Check-list and red-list of the liverworts (Marchantiophyta) and hornworts (Anthoeroceratophyta) of Italy. <i>Flora Mediterranea</i> 5:73–161.	Italy	Bryophytes	RL, CL
Conti F, Manzi A, Pedrotti F (1992) Libro rosso delle piante d'Italia. WWF Italia in collaborazione con la Società Botanica Italiana. (WWF, Camerino).	Italy	Vascular plants	RL, CL
Scientific Council of the Institute of Biology of Latvian Academy of Sciences (2000) Red Data Book of Latvia. <a href="http://enrin.grida.no/biodiv/biodiv/national/latvia/species/table_1.htm">http://enrin.grida.no/biodiv/biodiv/national/latvia/species/table_1.htm</a> .	Latvia	Various	RL, CL
Pilats V (1999) A short description of the Latvian mammal fauna. <i>Hystrix</i> 8:61–66.	Latvia	Mammals	RL, CL
Andrusaitis G, ed (2003) Red Data Book of Latvia. Rare and Threatened Species of Plants and Animals. Volume 5. Riga. Environment Protection Department of the Republic of Lithuania, ed (1992) Red Data Book of Lithuania.	Latvia	Reptiles	RL, CL
Balevicius K, Ladyga A (1992) Lietuvos raudonoji knyga [The Red Data Book of Lithuania]. (Lithuanian Department of Environmental Conservancy, Vilnius), 364 pp.	Lithuania	Various	RL, CL
Jukoniene I (1996) Rare and threatened bryophyte species in Lithuania. <i>Bot Lithuanica</i> 2:327–342.	Lithuania	Bryophytes	RL, CL
Creemers RCM, van Delft JJCW (2009) Rode Lijsten Amfibieën en Reptielen [Red List of reptiles and amphibians in the Netherlands]. (Nationaal Natuurhistorisch Museum Naturalis, European Invertebrate Survey, Leiden).	Netherlands	Reptiles	RL, CL
Siebel HN, Bijlsma RJ, Bal D (2006) Toelichting op de Rode Lijst Mossen. Rapport LNV Directie Kennis No. 2006/034.	Netherlands	Bryophytes	RL, CL
De Nie H (2003) Red listing of freshwater fishes and lampreys in the Netherlands. Available at <a href="http://home.planet.nl/~hwdenie/redlistfishes.pdf">http://home.planet.nl/~hwdenie/redlistfishes.pdf</a> .	Netherlands	Fish	RL, CL
Dutch Dragonflies (2011) Red list of Dutch Odonata 2011. Available at <a href="http://www.dutchdragonflies.eu/redlist.html">http://www.dutchdragonflies.eu/redlist.html</a> .	Netherlands	Dragonflies	RL, CL
Hustings F, Borgreve C, Van Turnhout C, Thissen J (2004) Basisrapport voor de Rode Lijst Vogels volgens Nederlandse en IUCN-criteria. SOVON onderzoeksrapport 2004/14. (SOVON Vogelonderzoek Nederland/Vogelbescherming Nederland, Beek-Ubbergen).	Netherlands	Birds	RL, CL
Kálás JA, Viken Á, Bakken T, eds (2006) Norsk Rødliste 2006 [2006 Norwegian Red List]. (Artsdatabanken, Norway).	Norway	Various	RL, CL
PAS (2004) Polish Red Data Book of Animals. Available at <a href="http://www.iop.krakow.pl/pckz/default.asp?nazwa=szukaj&amp;je=en">www.iop.krakow.pl/pckz/default.asp?nazwa=szukaj&amp;je=en</a> .	Poland	Animals	RL, CL
Zarzycki K, Wojewoda W, Heinrich Z (1992) Lista roślin zagrożonych w Polsce. 2nd ed. (Instytut Botaniki PAN, Kraków).	Poland	Bryophytes, vascular plants	RL, CL
Gowaqciniski Z, ed. (2002) Red list of threatened animals in Poland. (Polish Academy of Sciences, Institut of Nature Conservation).	Poland	Various	RL, CL
ICNB (2005) Red List of reptiles of Portugal. Available at <a href="http://www.icn.pt/destaques/destaques_anexos_L_Ver/repteis.pdf">www.icn.pt/destaques/destaques_anexos_L_Ver/repteis.pdf</a> .	Portugal	Reptiles	RL, CL
ICNB (2005) Red List of mammals of Portugal. Available at <a href="http://www.icn.pt/destaques/destaques_anexos_L_Ver/mamiferos.pdf">www.icn.pt/destaques/destaques_anexos_L_Ver/mamiferos.pdf</a> .	Portugal	Mammals	RL, CL
Almaça C (1995) Freshwater fish and their conservation in Portugal. <i>Biological Conservation</i> 72: 125–127.	Portugal	Fish	RL, CL
Oltean M, Negrean G, Popescu A, Roman N, Dihoru G, et al. (1994) Lista rosie a plantelor superioare din România [Red list of higher plants from Romania]. (Institutul de Biologie, Academia Româna), 52 pp.	Romania	Vascular plants	RL, CL
Botnariuc N, Tatole V, eds. (2005) Cartea Rosie a Vertebratelor din România [The Red Data Book of vertebrates of Romania]. (Academiei Române, Bucuresti).	Romania	Vertebrates	RL, CL
SOPSR (2010) Cervený zoznam ohrozených rastlín Slovenska. Available at <a href="http://www.sops.sk/webs/redlist/">www.sops.sk/webs/redlist/</a> .	Slovakia	Bryophytes, vascular plants	RL, CL
Balaz D, Marhold K, Urban P (2001) Cervený Zoznam Rastlín a Zvôcňoch Slovenska [Red Lists of Plants and animals of Slovakia]. <i>Ochrana Prírody Supplement</i> 20, 156 pp.	Slovakia	Various	RL, CL





**Table S3. Proportions of threatened species from seven taxonomic groups and 22 European countries as explained by multivariate models of historical or current socioeconomic indicators**

	1900	1950	2000
PCA 1	-0.29	0.19	-0.12
PCA 2	-0.15	-0.01	0.09
PCA 3	0.08	-0.17	-0.09
PCA 4	0.19	-0.06	-0.02
$R^2_{MF}$	0.22	0.11	0.05
AIC	4,807	5,539 <sub>ss</sub>	5,915
AW	>99.99	<0.001	<0.001

Fixed-effect estimates, McFadden's  $R^2$ , Akaike information criterion (AIC), and Akaike weights (AW) for generalized linear mixed-effects models relating the variation in the proportion of red-listed species to either historical (1900 and 1950) or contemporary (2000) values of impact indicators plus levels or recent expenditures on environmental protection. The three impact indicators—human population density (PD), per capita GDP (GDP), and human appropriation of net primary productivity (HANPP)—plus the environmental expenditures were subjected to a principal component analysis before fitting the regression models to remove multicollinearity among independent variables. All fixed-effect estimates are different from 0 with a probability > 0.999, except for the one of PCA 4 for the year 2000 ( $P = 0.95$ ).

**Table S4. Contemporary (year 2000) and historic (years 1900 and 1950) socioeconomic variables for the 22 European countries included in this study**

Country	PD <sub>1900</sub>	GDP <sub>1900</sub>	HANPP <sub>1900</sub>	PD <sub>1950</sub>	GDP <sub>1950</sub>	HANPP <sub>1950</sub>	PD <sub>2000</sub>	GDP <sub>2000</sub>	HANPP <sub>2000</sub>	EnvExp
Austria (AU)	71	2,882	0.291	83	3,706	0.237	98	20,097	0.342	0.726
Belgium (BE)	219	3,731	0.652	282	5,462	0.705	340	20,742	0.78	0.56
Bulgaria (BG)	36	1,223	0.296	65	1,651	0.384	70	5,365	0.274	0.367
Czech Republic (CZ)	110	2,400	0.388	113	3,501	0.330	130	9,047	0.406	0.43
Denmark (DK)	57	3,017	0.757	94	6,943	0.734	120	23,010	0.735	0.744
Finland (FI)	8	1,668	0.165	12	4,253	0.186	16	20,235	0.217	0.59
France (FR)	75	2,876	0.233	77	5,271	0.264	112	20,808	0.508	0.553
Germany (GE)	152	2,985	0.485	191	3,881	0.415	231	18,596	0.593	0.409
Greece (GR)	38	1,351	0.196	57	1,915	0.233	84	12,044	0.345	0.565
Hungary (HU)	77	1,682	0.434	100	2,480	0.342	109	7,138	0.525	0.575
Italy (IT)	134	1,785	0.233	187	3,502	0.331	231	18,740	0.406	0.859
Netherlands (NL)	143	3,424	0.449	282	5,996	0.709	454	21,591	0.924	1.7
Norway (NO)	7	1,937	0.060	10	5,463	0.067	14	24,364	0.058	0.657
Poland (PL)	79	1,536	0.315	80	2,447	0.284	124	7,215	0.377	0.525
Portugal (Pg)	61	1,302	0.156	95	2,086	0.118	118	14,022	0.293	0.516
Republic of Ireland (IR)	63	2,510	0.335	42	3,453	0.290	59	22,015	0.49	0.53
Romania (RO)	46	1,415	0.297	69	3,093	0.368	91	3,002	0.305	0.309
Slovakia (SK)	67	1,600	0.388	71	3,501	0.330	110	7,837	0.406	0.256
Spain (SP)	38	1,786	0.112	57	2,189	0.161	91	15,269	0.249	0.262
Sweden (SW)	11	2,561	0.180	16	6,739	0.145	20	20,321	0.174	0.299
Switzerland (SU)	80	3,833	0.242	114	9,064	0.323	176	22,025	0.381	0.85
United Kingdom (U.K.)	167	4,492	0.333	204	6,939	0.394	242	19,817	0.602	0.513

EnvExp, public expenditures in percent of GDP, means of the available data from the years 1998–2009; GDP, standardized per capita GDP in 1990 International (Geary–Khamis) dollars; HANPP, ratio of the amount of net primary productivity harvested by humans ( $NPP_h$ ) to the net primary productivity of a hypothetical undisturbed natural vegetation cover of the same area ( $NPP_0$ ); PD, human population density (people square kilometer).

